



KINETICS, ISOTHERMS AND THERMODYNAMICS STUDIES OF SORPTION OF Cu^{2+} ONTO NOVEL ZEROVALENT IRON NANOPARTICLES

By

A.O. Dada^{1*},

F.A. Adekola²

&

E.O. Odebumi³

^{1*}Department of Physical Sciences
(Industrial Chemistry Programme),
Landmark University,
Omu-Aran, Kwara State, Nigeria.

²Department of Industrial Chemistry,
University of Ilorin, Nigeria.

³Department of Chemistry,
University of Ilorin, Nigeria.

*Corresponding authors: dada.oluwasogo@lmu.edu.ng

Abstract: A novel nano-scaled zerovalent iron (nZVI) is an effective adsorbent for scavenging inorganic and organic toxicants. nZVI was synthesized in a single pot system using bottom-up approach and were characterized by BET, SEM, EDX and FTIR. In this study, sorption of Cu^{2+} onto nZVI was carried out vis-à-vis the investigation of physicochemical parameters (initial metal ion concentration, pH, temperature, adsorbent dose) at 298 K. The sorption data obtained at optimum conditions were subjected to six different isotherm models (Langmuir, Freundlich, Temkin, Dubinin-Raduskevich (DRK), Halsey and Harkin-Jura). However, the equilibrium sorption data were best described by both Langmuir and Temkin isotherm models with Langmuir maximum monolayer coverage (Q_{max}) of 40.816 mg/g and regression correlation value ($R^2 > 0.96$) supporting a chemisorption mechanism. Pseudo first-and second-order, Elovich, fractional power and intra-particle diffusion models were applied to the adsorption data in order to investigate the kinetic process; pseudo-second order fitted the data most. The intra-particle diffusion model suggested that the intra-particle diffusion was one of the rate-limiting steps. The values of the Gibbs free energy showed the feasibility and spontaneity of the sorption process. The

removal efficiency of Cu^{2+} (> 98%) onto zerovalent iron nanoparticles revealed that nZVI is a promising and efficient adsorbent that can be utilized by industries on a large scale for waste treatment.

Keywords: Zerovalent Iron nanoparticles; Sorption; Isotherms; Kinetics and Thermodynamics

Introduction

Copper is transition metal and one of the heavy metals in group 1B with density of 8.9 g.cm^{-3} . It is released into the environment through natural phenomena and anthropogenic activities and these account for its wide spread in the environment. Copper is often found near mines, industrial settings, landfills and waste disposals. Most copper compounds will settle and be bound to either water sediment or soil particles. Soluble copper compounds form the largest threat to human health. Usually, water-soluble copper compounds occur in the environment after release through application in agriculture. Copper in the blood exists in two forms either bound to *ceruloplasmin* or the rest "free", loosely bound to albumin and small molecules. Free copper causes toxicity, as it generates reactive oxygen species such as superoxide, hydrogen peroxide and the hydroxyl radicals all of which damage proteins, lipids and DNA [1]. There are a number of adverse effects of copper due to over-exposure such as: irritation of the nose, mouth and eyes, headaches, stomachaches, dizziness, vomiting, *hematemesis*, diarrhea, hypotension, melena, coma, jaundice. Intentionally, high uptakes of copper

may cause liver and kidney damage and even death [2].

Nanoparticles are the new trend of effective adsorbents used in the decontamination of water and immobilization of heavy metal ions from their solutions. Research on the utilization of nanoparticles is on the increase due to their special characteristics. Various methods have been reported by researchers for the removal of heavy metal ions from the environments such as chemical precipitation [3], ion exchange [4], solvent extraction [5], ultra-filtration [6], reverse osmosis [7], cementation [8], electro-dialysis [9]. These methods are not economical and have many disadvantages such as incomplete metal removal, high reagent and energy consumption, and generation of toxic sludge or other waste products that require disposal or treatment. In contrast, the adsorption technique is one of the preferred methods for the removal of heavy metal ions because of its efficiency and low cost [10].

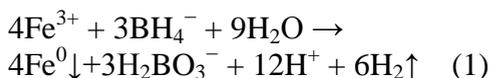
Although, previous researchers have investigated the removal of copper ions by utilizing natural clinoptilolite [11], amino-functionalized magnetic nanoparticles [12], pectin-iron oxide magnetic nanocomposite [13], Carboxymethyl-cyclodextrin conjugated magnetic nanoparticles

[14], polymer nanofibrous materials [15], titanate nanotubes [16], and palm shell activated carbon [17] yet to our knowledge, sorption of copper onto nZVI has not been the subject of detailed studies. The main objective of this research is to investigate the kinetics, isotherms and thermodynamics of sorption of copper ions onto core zerovalent iron nanoparticles vis-à-vis the optimization of sorption physicochemical parameters (pH, contact time, initial copper ions concentration and temperature).

Materials and Methods

Synthesis of Zerovalent Iron nanoparticles

All the chemicals used are of analytical grade purchased from Sigma Aldrich, USA. The two precursors used for the synthesis of zerovalent iron nanoparticles are 0.023 M solution of Ferric Chloride and 0.125 M solution of NaBH₄ in the ratio 1:5 respectively following the procedure reported elsewhere in the literature [18, 19]. For better formation of nano-scale zerovalent Iron (nZVI), more of NaBH₄ solution is needed than Fe³⁺ solution. As soon as the borohydride solution is added to Ferric chloride solution drop by drop in a glove box under inert environment. Ferric ion was reduced to zerovalent iron with an indication of a colour change according to the reaction:



The black particles of (nZVI) that appeared was further stirred for 3 h. nano-scale zerovalent Iron (nZVI) was separated from the solution using vacuum filtration apparatus and a cellulose nitrate membrane filter (Millipore filter) of 0.45 μm. nZVI was further washed with absolute ethanol three times and dried in a Genlab oven at 80 °C overnight.

SEM/EDX, FTIR and BET

Characterization of the adsorbent:

The zerovalent iron nanoparticles were characterized by a combination of scanning electron microscopy and energy dispersive x-ray. The scanning electron microscopy (SEM) was used for the determination of the structural morphology; the quantitative elemental determination was done through the energy dispersive x-ray. The FTIR spectrum provided information about the local molecular environment on the surface of nZVI. The analysis on the determination of surface area, pore size and volume were performed using Brunauer-Emmett-Teller (BET) and Barrett-Joyner-Halenda (BJH) methods.

Sorption Experiment

Batch Sorption Studies

The batch experiment was carried out following a similar procedure reported by Adekola *et al.* 2012 [20] and Dada *et al.* 2013 [21]. In a typical experiment, stock solution of 1000 ppm Cu²⁺ was prepared by dissolving a calculated amount of CuSO₄.5H₂O (3.927g) in 1000 cm³ of distilled deionized water in a standard

volumetric flask and serial dilution was made to prepare 10 – 100 ppm of the adsorbate. Batch sorption of Cu^{2+} ions onto nZVI was carried out and the sorption physicochemical parameters were determined. Generally, sorption experiment was done by contacting 100 mg of the nZVI with 50 cm^3 of different initial Cu^{2+} concentrations from 10ppm – 100 ppm in 60 cm^3 of Teflon bottle intermittently for 3 hours. The mixture was filtered and the filtrate was immediately analyzed for metal ions concentration using atomic absorption spectrophotometer (AAS) model AA320N at the Department of Physical Sciences, Industrial Chemistry Unit, Landmark University. The determination of the residual concentration using AAS was done in triplicate and the mean value for each set of the experiment was calculated. Optimization of sorption parameters (pH, contact time, adsorbent dose) was investigated following a similar procedure. Adsorption capacities were obtained using a mass balance equation [21]:

$$Q = \frac{(c_o - c_e)V}{m} \quad (2)$$

where Q is the equilibrium adsorption capacity per gram of nZVI (mg g^{-1}), V is the volume of Cu^{2+} solution (L), C_o is the initial concentration of the Cu^{2+} solution before sorption (mg L^{-1}), C_e is the final concentration of the Cu^{2+} solution after sorption (mg L^{-1}), m is the mass in gram of nZVI. The removal efficiency was calculated

using the formulae in Eq. (3) as reported in the literature [22]

$$\% RE = \frac{C_i - C_e}{C_i} \times 100 \quad (3)$$

The adsorption data obtained from the optimized parameters were tested and analyzed using kinetics models (pseudo-first and pseudo-second order, Elovich, power factor and intra-particle diffusion) and six different isotherm models (Langmuir, Freundlich, Tempkin, Dubinin-Raduskevich, Halsey and Harkin-Jura)

Optimization of pH, Adsorbent Dose, Contact time and Initial Cu^{2+} concentration

The effect of pH, adsorbent dose contact time, initial Cu^{2+} concentrations were investigated to determine the optimum conditions. The effect of pH on sorption of Cu^{2+} onto nZVI was done by adjusting the pH using 0.1 M HCl and 0.1 M NaOH solutions. The effect of pH on the sorption of Cu^{2+} was carried out within the pH range of 1 – 7 that will not influence Cu^{2+} precipitation [23, 24].

The effect of adsorbent dose was carried out at optimum pH, contact time and initial concentration of 100 mg L^{-1} and variable adsorbent doses of about 10 mg – 150 mg was contacted with 50 cm^3 portions of Cu^{2+} solution. The residual concentration was determined by AAS.

The effect of contact time on the sorption of Cu (II) ions onto nZVI was studied at various time intervals (0 – 120 min) at optimum pH and at different initial concentrations of 40 mgL⁻¹, 60 mgL⁻¹, 80 and 100mgL⁻¹. 100 mg adsorbent dose of nZVI was agitated with 50cm³ of the Cu(II) solution in 60 cm³ Teflon bottle at the optimum conditions and the residual Cu²⁺ concentration was determined in triplicate using AAS. The amount of metal ions sorbed was calculated using Eq. 2 and adsorption kinetics and mechanism were determined by data analysis using different kinetics models (pseudo first-order, pseudo second-order, Elovich, fractional power and intra-particle diffusion).

The effect of initial Cu²⁺ ions concentration on the extent of uptake was investigated as reported in the literature [25]. This was done by agitating 100 mg of nZVI added into 50 cm³ portions of Cu²⁺ solutions within the range of 10 – 100 mgL⁻¹ on the orbital shaker at optimum time of 60 minutes. The amount of metal ions sorbed was calculated and the data were fitted into six different isotherm models such as: Langmuir, Freundlich, Temkin, Dubinin-Raduskevich (DRK), Halsey and Harkin-Jura.

Results and Discussion

Characterization of Zerovalent Iron nanoparticles

Table 1 shows the physiochemical parameters of zerovalent iron nanoparticles. The Point of zero charge (PZC) is defined as the pH at

which the surface of the adsorbent (nZVI) has a net neutral charge. The PZC of nZVI as determined by salt addition method is 5.28 (Fig not shown). The significance of this is that nZVI has a positive charge at solution pH values less than the PZC and thus a surface on which anions may adsorb. On the other hand, nZVI has a negative charge when the pH of the solution is greater than the PZC and thus be a surface on which cation may adsorb and thus the pH of nZVI is 5.28 indicating that nZVI better adsorption capacity was obtained at as solution pH greater than 5.28 stated on Table 1 [26].

The surface area by BET is 20.8643 m²/g, the t-Plot micropore area is 4.4140 m²/g which is lower than the t-Plot external surface area (is 16.4503 m²/g). These results revealed that most of the Cu²⁺ ions are adsorbed at the external surface area and hence adsorption prevails. The large surface of nano-scaled zerovalent iron (nZVI) enhances its relevance in the waste water remediation [27].

The SEM micrograph of ZVI showed a spherical surface of chain-like, aggregated molecules. The chain-like aggregation is an indication of its magnetic property. The EDX spectrum of nZVI (Fig 1) showed the characteristics peaks of the nZVI, the information on the surface atomic distribution and the chemical elemental composition couple with the percentage composition of the prepared are stated on table 2. From table 2, iron (Fe) has the highest

atomic weight and atomic abundance of 86.47% and 51.28% respectively, this is in accordance with the report of [28, 29].

Figure 2 showed the FTIR spectrum of nZVI. The peaks on Fig. 2 are 3345 cm^{-1} , 3256 cm^{-1} , 1632 cm^{-1} , 1328 cm^{-1} , 910 cm^{-1} , and 686 cm^{-1} . The broad and intense peak around 3345 cm^{-1} , 3256 cm^{-1} is due to the presence of O—H from alcohol used in washing nZVI to prevent rapid corrosion. The peak at 1632 cm^{-1} may be attributed to H—O—H stretching of deionized deoxygenated water, 1328 cm^{-1} corresponds to C—H bending, 910 cm^{-1} is due to C—H bending out of plane and 686 cm^{-1} is attributed to zerovalent iron, Fe^0 as reported in the literature [28, 29, 30].

Optimization of pH, Contact time and Initial Concentration.

Effect of pH places one of the greatest roles in the adsorption studies because it influences the surface charge of the adsorbents, ionic mobility, the degree of ionization and speciation of different pollutants and solution chemistry of contaminants (i.e. hydrolysis, redox reactions, polymerization and coordination) [26]. Fig 3 shows the plot the effect of pH on the sorption of Cu^{2+} onto nZVI. The percentage removal efficiency increases greatly from pH 4 (68.158 %) to pH 6 (86.3 %) where it attained equilibrium. The low percentage removal efficiency observed from pH 1 to 3 is a result of increase in electrostatic effect arising from the competition among Cu^{2+} , H_3O^+ and

other cationic species ($\text{Cu}(\text{OH})^+$, $\text{Cu}(\text{OH})_2$, $\text{Cu}(\text{OH})_3^-$ and $\text{Cu}(\text{OH})_4^{2-}$) for adsorption sites at lower pH. As the pH of the solution increases, the competition for adsorption site reduces and more of Cu^{2+} dominates thereby leading to increase in the quantity adsorbed (34.08 mg/g – 43.15 mg/g) and considerable increase in the percentage removal efficiency. The optimum pH observed was pH 6 and all other optimization studies were carried out at this pH [12, 13, 14, 26].

The effect of adsorbent dose on the sorption of Cu^{2+} onto nZVI is shown on Fig 4. Increase in the adsorbent dose from 18.99 mg/g at 10mg to 22.62 mg/g was a result of increase in the number of active site as the adsorbent increases. This is similar to findings reported in the literature [30].

The effect of contact time (Fig 5) is another important factor in all transfer phenomena such as adsorption. A short contact time to reach equilibrium indicates the fast transport of Cu^{2+} ions from the bulk to the outer and inner surface of nZVMn. In addition, contact time also controls the buildup of charges at the solid-liquid interfaces and for this reason, optimization of the contact time on the effect of sorption of Cu^{2+} onto nZVI was investigated at three different initial concentrations 40ppm, 60ppm and 100ppm from 10min to 120min. The rate of reaction was rapid from 10 min with 16.878 mg/g, 25.854 mg/g and 40.03 mg/g

(Fig 5) at the three concentrations respectively. The contact time of 60 minutes was observed as the optimum time after which a steady state approximation set in and a quasi-equilibrium situation was attained. All other optimization studies were carried out at 60 minutes contact time. Fig. 6 shows the result of the effect of initial Cu^{2+} ion concentration on the sorption of Cu^{2+} onto nZVI. As the initial concentration increased from 10ppm to 100ppm, the percentage removal efficiency also increased as a result of available and vacant active sites at low concentration and as the concentration increased, the active site became saturated and a state of equilibrium was attained where there was no significant increase in the percentage removal efficiency [25, 31]

Adsorption Kinetics

The kinetic models were exploited to test the experimental data obtained from contact times ranging from 10 to 120 minutes by monitoring the quantity of Cu^{2+} adsorbed. In order to determine the mechanisms of sorption and its potential rate controlling steps which include mass transport and chemical reaction process, pseudo first-order, pseudo second-order, Elovich, power function (fractional power) and intraparticle diffusion rate equations have been used to model the kinetics of sorption of Cu^{2+} onto nZVI. The constant parameters, correlation coefficient (R^2) and their values are summarized on Table 3

The pseudo first-order equation (Lagergren's equation) describes adsorption in solid-liquid systems based on the sorption capacity of solids. It is assumed that one copper ion was sorbed onto one sorption site on the nZVI surface [32]:



Where S represents an unoccupied sorption site on the nZVI
The pseudo first-order equation is generally expressed as:

$$\frac{dq}{dt} = k_1(q_e - q_t) \quad (5)$$

where q_e is the amount of Cu^{2+} adsorbed at equilibrium per unit weight of the adsorbent (mg/g), q_t is the amount of Cu^{2+} adsorbed at any time (mg/g) and k_1 is the pseudo first-order rate (constant/min). After the equation is integrated and boundary conditions ($t = 0 \rightarrow t$ and $q_t = 0 \rightarrow q_t$) are applied the equation becomes:

$$\text{Log}(q_e - q_t) = \text{Log } q_e - \frac{k_1 t}{2.303} \quad (6)$$

The values of $\text{Log}(q_e - q_t)$ were linearly correlated with t . The plot of $\text{Log}(q_e - q_t)$ versus t as shown on Fig. 7 gave a linear relationship and k_1 and q_e were determined from the slope and intercept of the expression in Eq. 5 respectively. The parameters are summarized on Table 3 and it showed that the kinetics model did not fit into pseudo first order model equation because the correlation

coefficient was very low and the adsorption capacity calculated using this model was not close to the experimental quantity adsorbed indicating that the data were poorly fitted into pseudo first order (Hao et al., 2010)

The pseudo second-order rate expression, which has been applied for analyzing Chemisorption kinetics from liquid solutions [31, 32].

$$\frac{dq}{dt} = k_2(q_e - q_t)^2 \quad (7).$$

where k_2 is the rate constant of for pseudo second-order adsorption equation (g/mg min). For the boundary conditions ($t = 0 \rightarrow t$ and $q_t = 0 \rightarrow q_t$), the integrated form of Eq. (6) is linearly expressed as:

$$\frac{t}{q_t} = \frac{1}{k_2 q_e^2} + \frac{1}{q_e} t \quad (8)$$

$$h_0 = k_2 q_e^2 \quad (9)$$

Substituting h_0 into equation above, it becomes:

$$\frac{t}{q_t} = \frac{1}{h_0} + \frac{1}{q_e} t \quad (10)$$

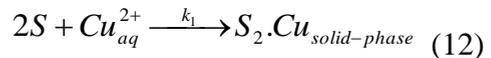
where h_0 is the initial adsorption rate. The plot of t/q_t against t as shown on Fig. 8 gave a linear plot with good and commendable correlation coefficients close to unity at each concentration as shown on Table 3. The pseudo second-order parameters,

k_2 , q_e and h_0 were determined from the slope and intercept of Eq. 10

When t tends to 0, h_0 is defined as:

$$h_0 = k_2 q_e^2 \quad (11)$$

This model assumes that one copper ion is sorbed onto two sorption sites on the nZVI surface:



From the parameters summarized on Table 3, the adsorption kinetics can be best described by pseudo second-order model. The R^2 values at each concentration are greater than 0.99 and the calculated quantity adsorbed was perfectly close to experimental quantity adsorbed. This is in accordance with the findings reported in the literature [18, 32, 33, 34]

The Elovich model is generally described as:

$$\frac{dq}{dt} = \alpha \exp(-\beta q_t) \quad (13)$$

Applying the boundary conditions ($q_t=0$ at $t = 0$ and $q_t = q_t$ at $t = t$), the simplified form of the Elovich equation is expressed as [35]:

$$q_t = \frac{1}{\beta} \ln\left(\frac{\alpha}{\beta}\right) + \frac{1}{\beta} \ln(t) \quad (14)$$

where q_t is the amount of adsorbate per unit mass of sorbent at time (t). The initial adsorption rate, α (mg/g-min) and the desorption constant (β)

(g/mg) during any experiment were determined from the slope($1/\beta$) and intercept ($(1/\beta)\ln(\alpha/\beta)$) of linear plot of q_t versus (t) as shown on Fig. 9. From the parameters on Table 3, it can be deduced that since the values of R^2 were to unity and the adsorption rates were also high, the data could also be described by the Elovich model.

The Fractional power model can be expressed as [20,31]:

$$q_t = kt^v \quad (15)$$

The equation above is linearized as:

$$\log(q_t) = \log(k) + v \log(t) \quad (16)$$

where q_t is the amount of sorbate per unit mass of sorbent, k is the fractional power rate constant, t is time, and v is a positive constant (<1). The parameters v and k were determined from slope and intercept of a linear plot of $\log(q_t)$ versus $\log(t)$ as shown on Fig. 10. The parameters shown on Table 3 indicated that the parameter fit also into the fractional model because the values of v were less than 1 and the regression coefficient (R^2) was greater than 0.96 at optimum concentration [30]

The intraparticle diffusion equation as expressed by Weber-Morris is given as [35]:

$$q_t = k_{id}t^{0.5} + C \quad (17)$$

Where k_1 is the intraparticle diffusion rate constant ($\text{mg g}^{-1} \text{min}^{1/2}$) and C is the intercept determined from the linear plot of q_t versus $t^{0.5}$ (Fig. 11). The intercept of the plot reflects the boundary layer effect. Higher values of the boundary layer effect obtained (16.246, 25.628 and 39.667) are indication greater contribution of the surface sorption in the rate controlling step. The calculated intraparticle diffusion coefficient (k_{id}) values are listed in Table 3. Since the linear plot of q_t versus $t^{1/2}$ did not pass through the origin, then intraparticle diffusion not the sole rate-limiting step [36].

Adsorption Isotherms

The adsorption isotherm constant parameters, correlation coefficient (R^2) and their values are summarized in Table 4.

Langmuir Adsorption Isotherm assumes a monolayer adsorption onto a homogeneous surface with a finite number of identical sites also there is uniform energy of adsorption onto the surface and no transmigration of adsorbate in the plane of the surface [37, 38]. The linear form of Langmuir represents:

$$\frac{C_e}{Q_e} = \frac{1}{K_L Q_{\max}} + \frac{C_e}{Q_{\max}} \quad (18)$$

Q_{\max} is the maximum monolayer coverage capacity (mg.g^{-1}), K_L is the Langmuir isotherm constant (L.mg^{-1}) related to the energy of adsorption. The essential features of the

Langmuir isotherm may be expressed in terms of equilibrium parameter R_L , which is a dimensionless constant referred to as separation factor or equilibrium parameter [38].

$$R_L = \frac{1}{1 + K_L C_o} \quad (19)$$

Figure 12 shows the Langmuir isotherm model plot for sorption of Cu^{2+} onto nZVI. The Langmuir constants, Q_{\max} and K_L were determined from the linear plot of C_e/Q_e versus C_e as shown on Fig 12. The value of separation factor, R_L (Fig 13) indicates the adsorption nature to either unfavourable or unfavourable. If $R_L > 1$, it is unfavorable, linear if $R_L = 1$, favorable if $0 < R_L < 1$ and irreversible if $R_L = 0$. Since the value of separation factor (R_L) ranges from 0.110 at 10ppm to 0.021 at 100ppm and none of them was greater than 1 therefore, the adsorption of Cu^{2+} onto nZVI was favourable. In addition, the regression coefficient ($R^2 > 0.90$) and maximum monolayer coverage capacity ($Q_{\max} = 40.816 \text{ mg/g}$) revealed that the equilibrium sorption can be best described by Langmuir model which supported a chemisorption mechanism

The Freundlich sorption isotherm gives an expression encompassing the surface heterogeneity and the exponential distribution of active sites

and their energies. The Linear form of Freundlich equation is [36, 37]:

$$\log Q_e = \log K_f + \frac{1}{n} \log C_e \quad (20)$$

The Freundlich isotherm constants, K_f and n are parameters characteristic of the sorbent-sorbate system, which were determined from the intercept and slope of the plot of $\text{Log}Q_e$ against $\text{Log}C_e$ (Fig 14) from the Table 4, the value of n (2.853) less than 10 is also an indication of a favourable adsorption. However, the sorption data could not be best described by Freundlich model compared to Langmuir model.

Temkin Isotherm contains a factor that explicitly taking into the account of adsorbent–adsorbate interactions in a chemisorption mechanism. The model assumes that heat of adsorption (function of temperature) of all molecules in the layer would decrease linearly with the surface coverage due to adsorbent–adsorbate interactions. The linear form of the equation is given as [39]:

$$Q_e = \frac{RT}{b_T} \ln A_T + \frac{RT}{b_T} \ln C_e \quad (21)$$

Where $B = RT/b_T$, b_T is the Temkin isotherm constant related to the heat of sorption and A_T is the Temkin isotherm equilibrium binding constant (Lg^{-1}). The values of these constants were determined from the slope and intercept obtained from appropriate plot of Q_e versus $\ln C_e$ as shown in figure 15 and are clearly shown on

Table 4. The high value of coefficient of regression ($R^2 = 0.937$) obtained from Temkin model corroborated the chemisorption mechanism as inferred from Langmuir Model.

Dubinín–Kaganer–Radushkevich (DKR) isotherm model is generally applied to express the adsorption mechanism with a Gaussian energy distribution onto a heterogeneous surface. The model has often successfully fitted high solute activities and the intermediate range of concentrations data well. The linear equation is given by [36, 37, 38, 39]:

$$\ln Q_e = \ln Q_d - A_{D-R} \varepsilon^2 \quad (22)$$

Where A_{D-R} is the DRK isotherm constant (mol^2/kJ^2) related to free sorption energy and Q_d is the theoretical isotherm saturation capacity (mg/g). The values of A_{D-R} and Q_d were determined respectively from the slope and intercept of the plot of $\ln Q_e$ versus ε^2 (Fig 16). The parameter ε is the Polanyi potential which is computed as:

$$\varepsilon = RT \ln \left[1 + \frac{1}{C_e} \right] \quad (23)$$

The approach was usually applied to distinguish the physical and chemical adsorption of metal ions with its mean sorption free energy, E per molecule of adsorbate (for removing a molecule from its location in the sorption space to the infinity) can be computed by the relationship [40]:

$$E = - \left[\frac{1}{\sqrt{2A_{D-R}}} \right] \quad (24)$$

If the value of mean sorption free energy (E) is less than 8kJ mol^{-1} , the adsorption mechanism will be physisorption and if the value of E is greater than 8kJ mol^{-1} , the adsorption will be chemisorption. Since the magnitude of E (free energy of transfer of one solute from infinity to the surface of nZVI) is greater than 8KJ mol^{-1} (Table 4), the adsorption mechanism is chemisorption which further supported Langmuir Isotherm [39].

Both Halsey isotherm and Harkin-Jura are used to evaluate the multilayer adsorption at a relatively large distance from the surface. The Halsey isotherm is expressed by [36, 39]:

$$\ln q_e = \left[\left(\frac{1}{n_H} \right) \ln K \right] - \left(\frac{1}{n_H} \right) \ln C_e \quad (25)$$

A plot of $\ln q_e$ against $\ln C_e$ gave a linear graph (Fig 17) and the Halsey constants K_H and n_H were determined from the intercepts and slope respectively. The Halsey isotherm model is always in support of Freundlich model and an evidence of this is shown from the regression coefficient ($R^2 = 0.874$) which is similar to that obtained from Freundlich model. Both Freundlich and Halsey models always support

multilayer adsorption [39]. A low regression coefficient obtained from both Freundlich and Hasley models is an indication that the equilibrium data did not follow multilayer adsorption.

The Harkin-Jura isotherm is expressed as:

$$\frac{1}{q_e^2} = \frac{B_{HJ}}{A_{HJ}} - \frac{1}{A_{HJ}} \log Ce \quad (26)$$

The Harkin-Jura constants A_{HJ} and B_{HJ} were determined from the slope and intercept of the linear plot of $\frac{1}{q_e^2}$ versus $\log Ce$ as shown on figure 18. The parameters of Harkin-Jura isotherm model are as shown on Table 4 revealed that the equilibrium sorption data is poorly described by this model which further supported chemisorption process.

Thermodynamics studies

Temperature is another important parameter in the adsorption studies because some important parameters like enthalpy change (ΔH), entropy change (ΔS) and Gibbs free energy change (ΔG) could be determined. The thermodynamic parameters can be determined from the thermodynamic equilibrium constant, K_a (or the thermodynamic distribution coefficient). The standard Gibbs free energy ΔG° (kJ mol^{-1}), standard enthalpy change ΔH° (kJ mol^{-1}), and standard entropy change ΔS° ($\text{J mol}^{-1}\text{K}^{-1}$) were calculated using the following equations [30, 40]:

$$Ka = \frac{q_e}{Ce} \quad (27)$$

$$\Delta G = -RT \ln Ka \quad (28)$$

$$\text{Log}Ka = \frac{\Delta S}{2.303R} - \frac{\Delta H}{2.303RT} \quad (29)$$

The Van't Hoff plot on sorption of Cu^{2+} onto nZVI is shown on Fig. 19 and the thermodynamics parameters are stated on Table 5. The positive standard enthalpy change ($\Delta H^\circ = 53.44925 \text{ kJmol}^{-1}$) for this study suggests that the adsorption of Cu^{2+} by nZVI is endothermic in nature, which is supported by the increase in Cu^{2+} adsorption with increase in temperature from 298 K to 318 K. The positive standard entropy change ($\Delta S^\circ = 45.3 \text{ Jmol}^{-1} \text{ K}^{-1}$) reflected the affinity of the nZVI particles towards Cu^{2+} and the negative values of the standard Gibbs free energy ΔG° (Table 5) is an indication of the feasibility and spontaneity of the adsorption process.

Conclusion

The synthesis of nanoscaled zerovalent iron via single pot system vis-à-vis chemical reduction is a novel approach. The isotherms, kinetics and thermodynamics studies of sorption of Cu^{2+} onto zerovalent iron nanoparticles was investigated. Optimization studies revealed that pH, contact time, initial concentration, adsorbent dose and temperature played substantial role in

adsorption studies and maximum adsorption capacity can be obtained at pH 6 for sorption of Cu^{2+} . The pseudo first-order, pseudo second-order, Elovich, power function (fractional power) and intraparticle diffusion rate equations have been used to model the kinetics of sorption of Cu^{2+} and from this investigation, pseudo second-order best described the kinetic process. From the six different adsorption isotherm models investigated, equilibrium sorption data fitted best into Langmuir and Temkin isotherm models indicating a Chemisorption mechanism and this was also supported with the high energy value calculated from DRK isotherm model which is in the range of the energy values for

Chemisorption process. The values obtained from the thermodynamics parameters, ΔH , ΔS , and ΔG vividly proved the reaction was feasible, spontaneous and endothermic in nature. Outcome of this research showed that zerovalent iron nanoparticle is an effective and promising nano-adsorbent for the sorption of toxic heavy metal ions.

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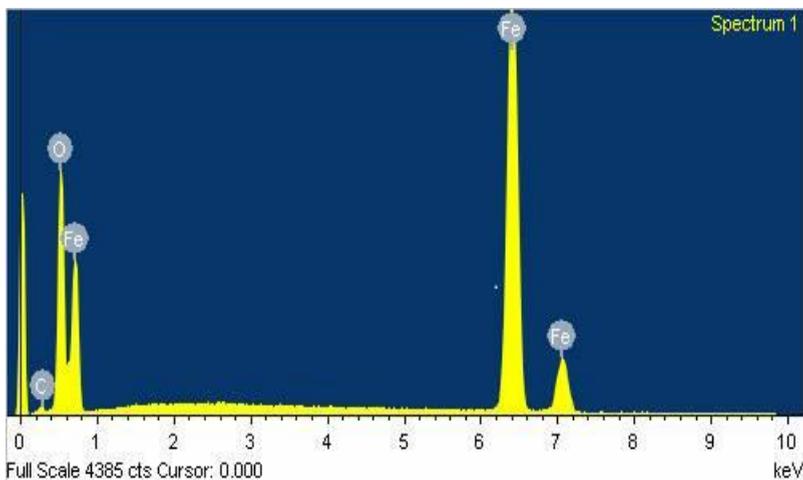
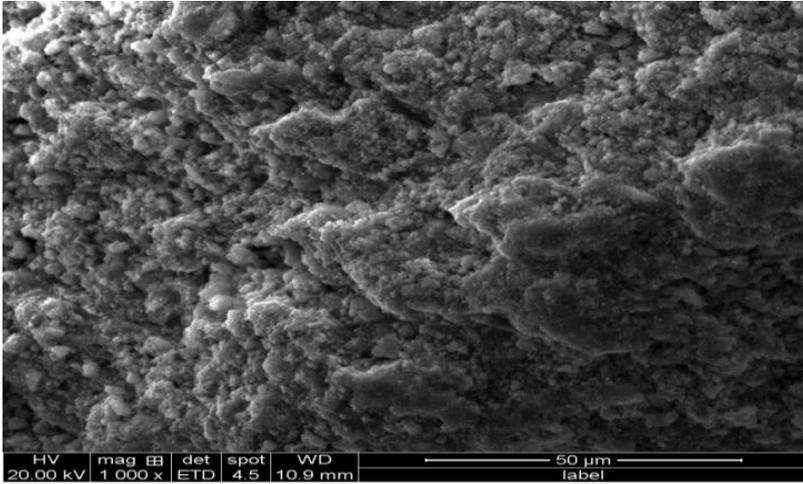


Fig 1: EDX spectrum for nZVI



SEM micrograph for nZVI

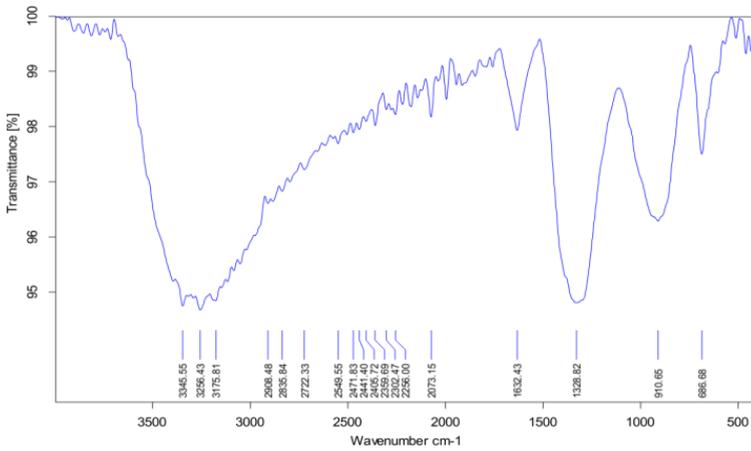


Fig 2: FTIR spectrum for nZVI

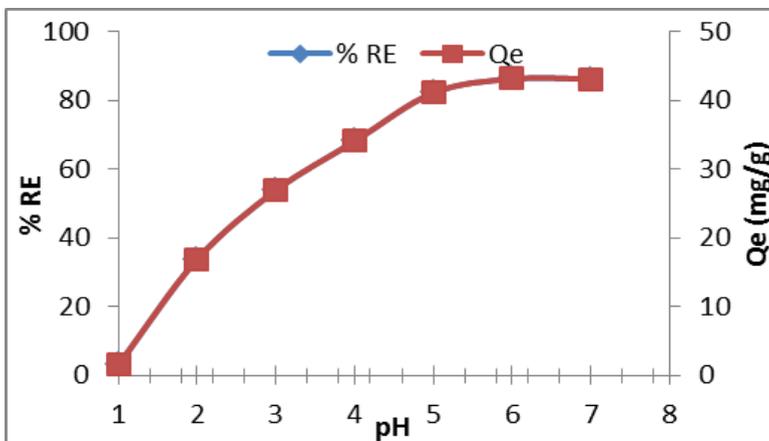


Fig 3: Effect of pH on sorption of Cu²⁺ onto nZVI

Experimental conditions: Cu²⁺ Concentration = 100 mg/L
 Volume of Cu²⁺ solution = 50 mL; Contact time = 60 min,
 Stirring speed = 200 rpm, Adsorbent Dose = 100 mg
 and Temperature = 25 ± 2 °C

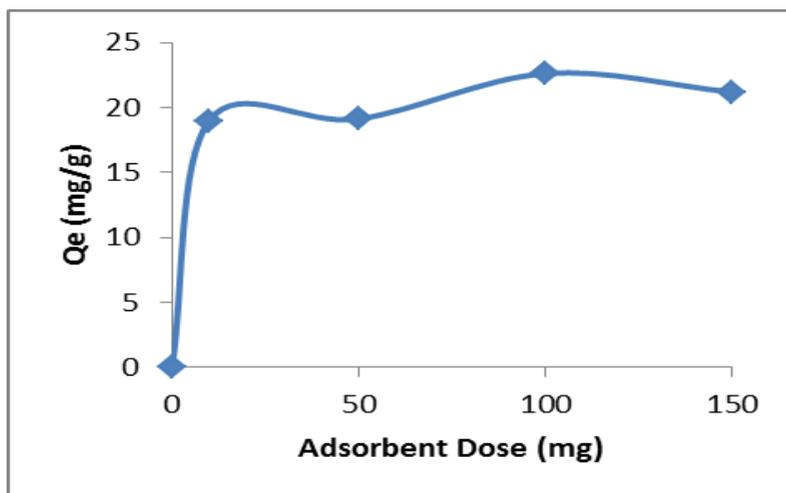


Fig 4: Effect of Adsorbent dose on sorption of Cu²⁺ onto nZVI

Experimental conditions: Cu²⁺ Concentration = 100 mg/L
 Volume of Cu²⁺ solution = 50 mL; pH = 6, stirring speed = 200 rpm,
 Contact time = 60 min, and temperature = 25 ± 2 °C

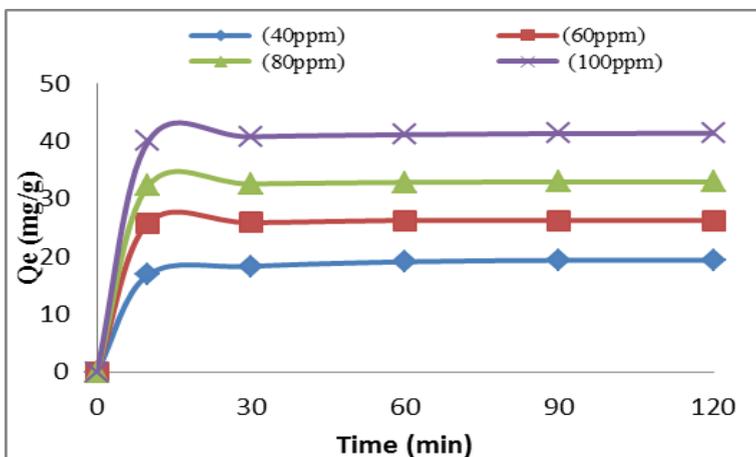


Fig 5: Effect of Contact time on sorption of Cu²⁺ onto nZVI

*Experimental conditions: Cu²⁺ Concentration= 100 mg/L
 Volume of Cu²⁺ solution = 50 mL; pH = 6, stirring speed = 200 rpm
 Adsorbent Dose =100 mg and temperature = 25 ± 2 °C*

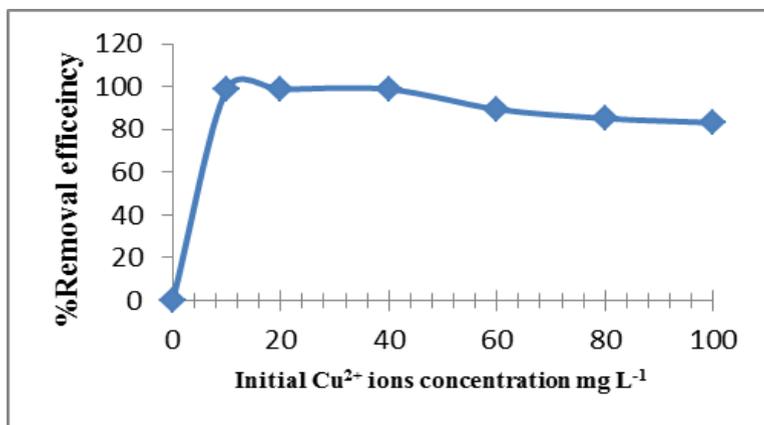


Fig 6: Effect of Initial Concentrations

*Experimental conditions: Volume of Cu²⁺ solution = 50 mL;
 Contact time = 60 min, pH = 6; Stirring speed = 200 rpm;
 Adsorbent Dose =100 mg and Temperature = 25 ± 2 °C*

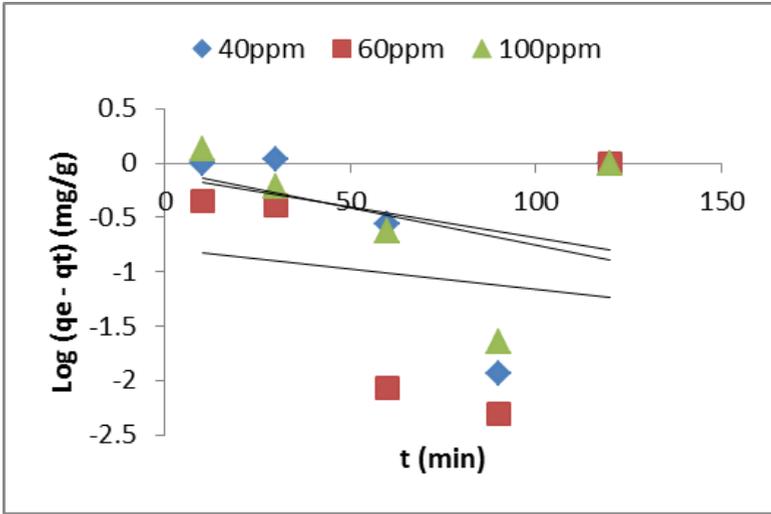


Fig 7: Pseudo first-order kinetic of sorption of Cu^{2+} onto nZVI

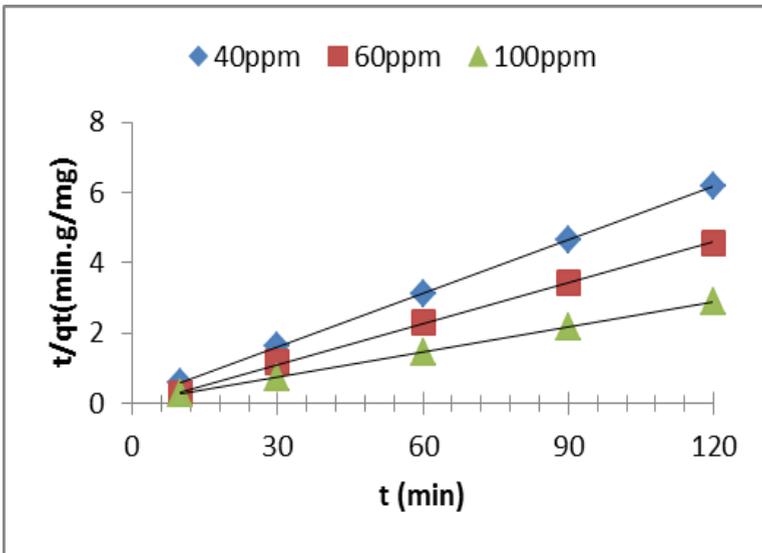


Fig 8: Pseudo second-order kinetic of sorption of Cu^{2+} onto nZVI

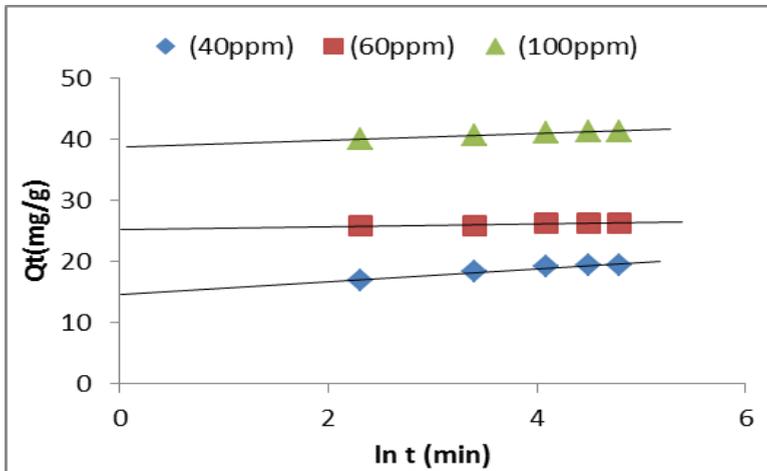


Fig 9: Elovich model of sorption of Cu^{2+} onto nZVI

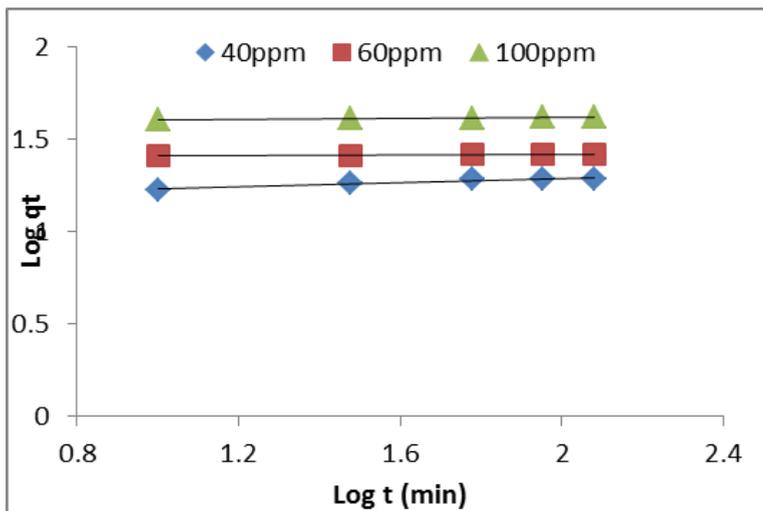


Fig 10: Fractional power of sorption of Cu^{2+} onto nZVI

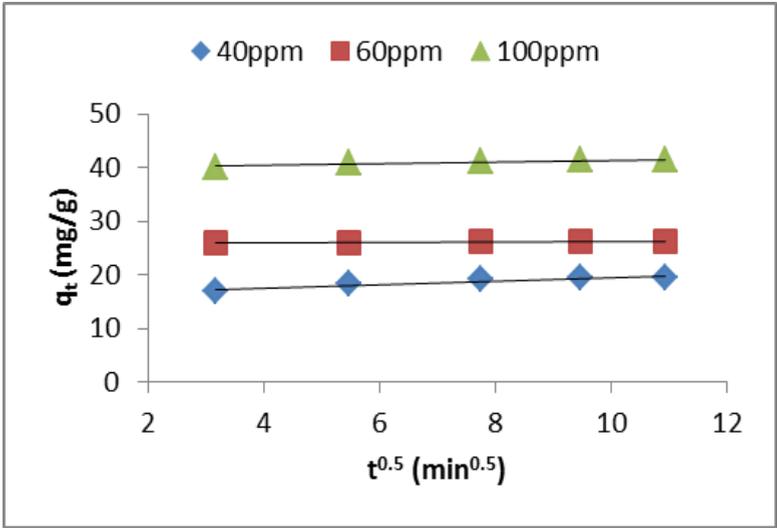


Fig 11: Intraparticle diffusion plot for sorption of Cu^{2+} onto nZVI

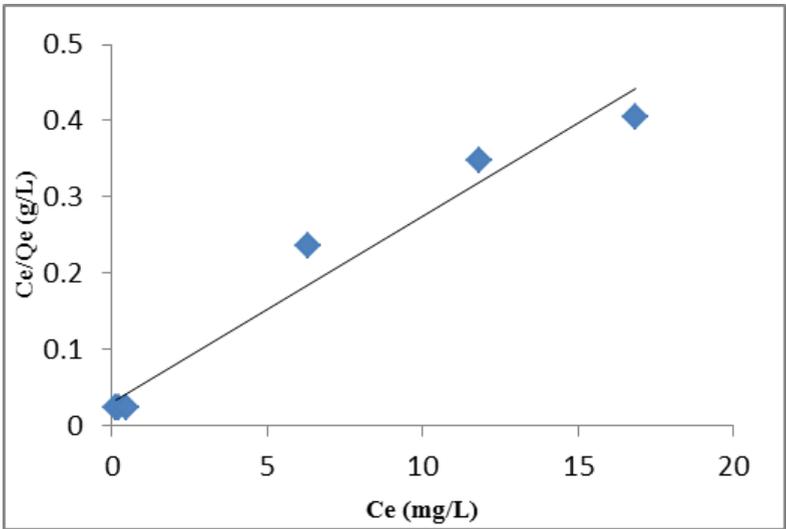


Fig 12: Langmuir Isotherm model for sorption of Cu^{2+} onto nZVI

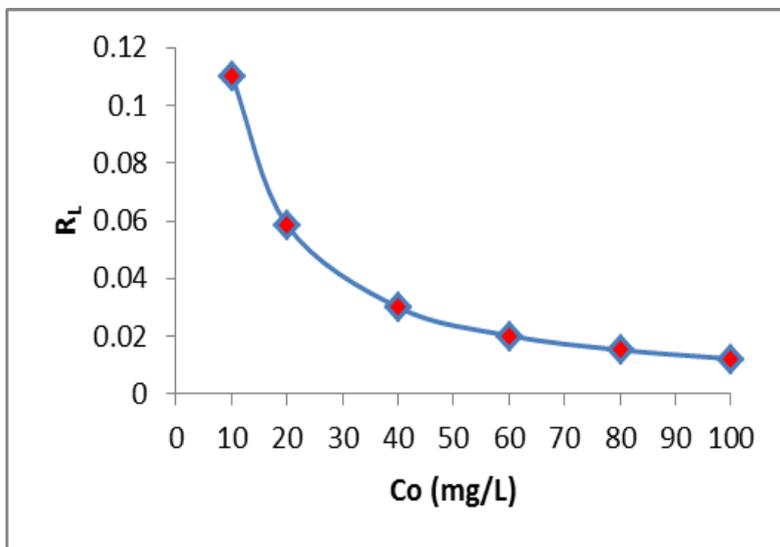


Fig 13: Separation factor on sorption of Cu^{2+} onto nZVI

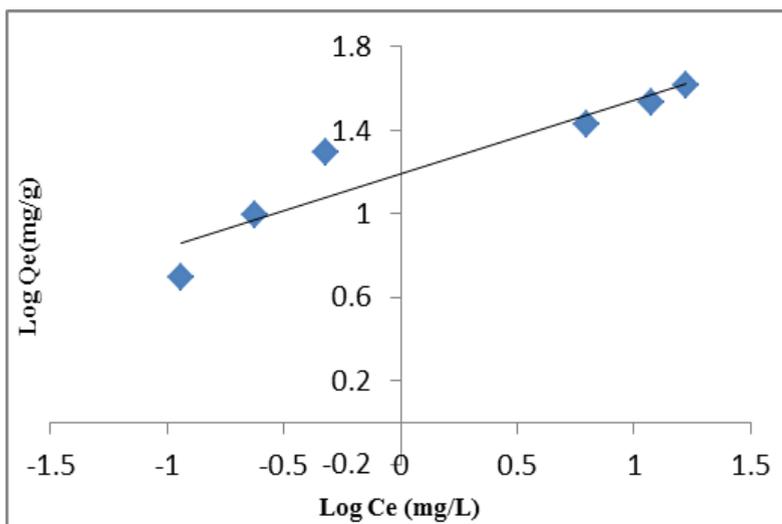


Fig 14: Freundlich Isotherm model for sorption of Cu^{2+} onto nZVI

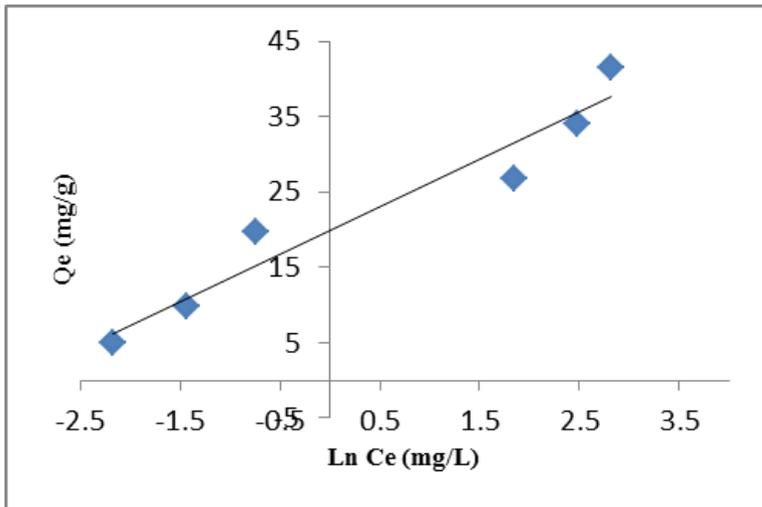


Fig 15: Temkin Isotherm model for sorption of Cu^{2+} onto nZVI

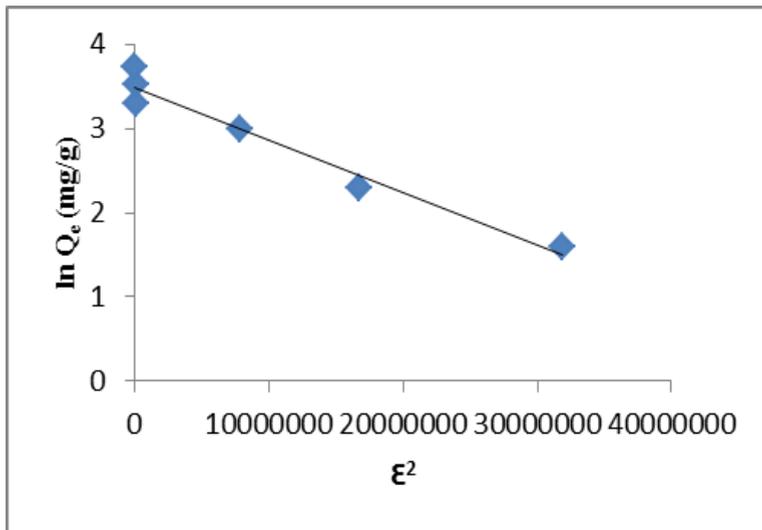


Fig 16: DRK Isotherm model for sorption of Cu^{2+} onto nZVI

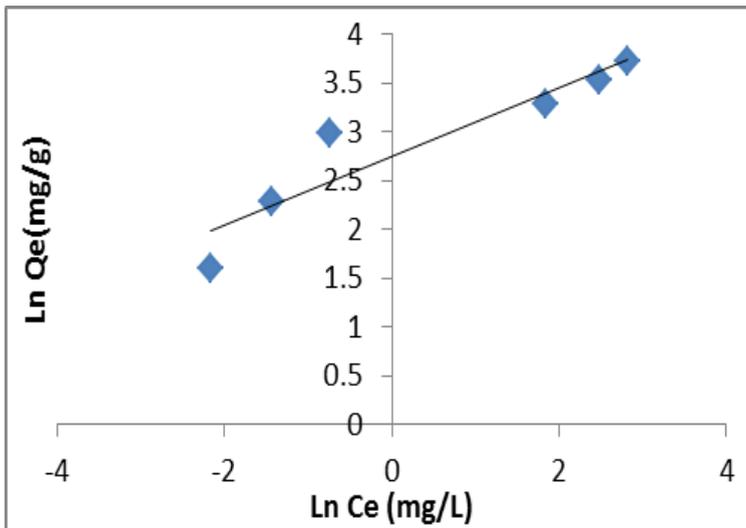


Fig 17: Halsey Isotherm model for sorption of Cu^{2+} onto nZVI

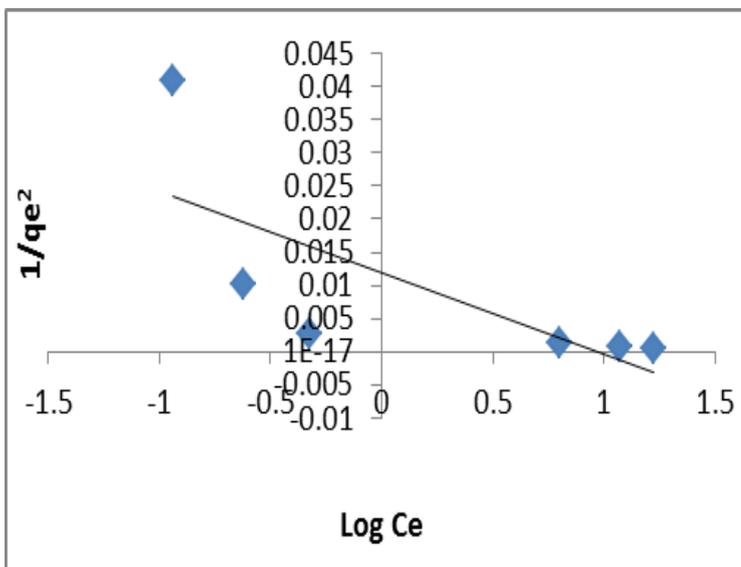


Fig 18: Harkins-Jura Isotherm model for sorption of Cu^{2+} onto nZVI

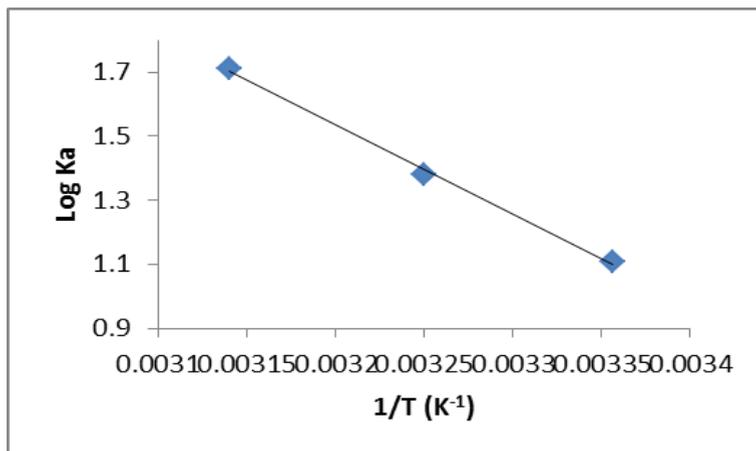


Fig 19:Van't Hoff plot on sorption of Cu²⁺ onto nZVI

Table 1: Physicochemical parameters of nZVI as adsorbent for the sorption of Cu²⁺

pH	6.00
PZC	5.20
BET Surface Area:	20.8643 m ² /g
t-Plot Micropore Area:	4.4140 m ² /g
t-Plot External Surface Area:	16.4503 m ² /g
BJH Adsorption cumulative surface area of pores between 17.000 Å and 3000.000 Å diameter:	19.120 m ² /g
Pore Volume	
Single point adsorption total pore volume of pores less than 1070.322 Å diameter at P/Po = 0.981571614:	0.097502 cm ³ /g
t-Plot micropore volume:	0.001895 cm ³ /g
BJH Adsorption cumulative volume of pores between 17.000 Å and 3000.000 Å diameter:	0.115083 cm ³ /g
Pore Size	
Adsorption average pore width (4V/A by BET):	186.9268 Å
BJH Adsorption average pore diameter (4V/A):	240.753 Å

Table 2: EDX Elemental percentage composition of NZVI

Element	Weight%	Atomic%
C K	1.72	4.75
O K	21.24	43.97
Fe K	86.47	51.28
Totals	109.43	100

Table 3: Kinetic model parameters for Cu²⁺ adsorption onto nZVI investigated at different concentrations

Models	40ppm	60ppm	.00ppm
pseudo first-order			
k_1	0.0159	0.00875	0.0131
qe,exp	19.388	26.305	41.39
qe,cal	0.855	0.163	0.763
R^2	0.131	0.0242	0.124
Pseudo second-order			
k_2	0.0264	0.1177	0.0538
qe,exp	19.388	26.305	41.39
qe,cal	19.723	25.773	41.49
H_0	10.256	13.928	92.593
R^2	1	0.999	1
Elovich			
α	10.09×10^5	48.196	3.932
β	0.944	4.673	1.771
R^2	0.968	0.837	0.982
Fractional Power			
K	14.887	25.322	38.833
V	0.0583	0.0082	0.0139
R^2	0.963	0.837	0.981
Intra-particle Diffusion			
K	0.0322	0.0683	0.1735
C	16.246	25.628	39.667
R^2	0.881	0.835	0.909

Table 4: Langmuir, Freundlich, Temkin, and D–R, Halsey and Harkin-Jura isotherm models parameters and correlation coefficients for adsorption of copper ions onto nZVI particles

Isotherm Models	Parameters	Cu ²⁺
Langmuir	$q_{max} (mg g^{-1})$	40.816
	$K_L (Lmg^{-1})$	0.806
	R_L	0.012
	R^2	0.966
Freundlich	k_f	15.488
	$1/n$	0.351
	n	2.853
	R^2	0.874
Temkin	$b_T (J mol^{-1})$	4.501×10^{-8}
	$B (Lg^{-1})$	6.2863
	$A_T (Lg^{-1})$	4.501×10^8
	R^2	0.937
DRK	q_d	32.769
	A_{DRK}	-6.008
	$E(KJ/mol)$	9.128
	R^2	0.961
Halsey	$1/n$	0.3505
	N	2.853
	K	2484.157
	R^2	0.874
Harkin-Jura	A	81.3008
	B	0.9675
	R^2	0.533

Table 5: Thermodynamic parameters for adsorption of Cu²⁺ onto nZVI

T (K)	$\Delta G(\text{kJ mol}^{-1})$	$\Delta H (\text{kJ mol}^{-1})$	$\Delta S (\text{J mol}^{-1} \text{ K}^{-1})$	Ka
298	6.31409	53.44925	200.4323	12.782
308	8.146.56			24.065
318	10.4076			51.204